# Mesohabitat Use and Community Structure of Brazos River Fishes in the Vicinity of the Proposed Allens Creek Reservoir 

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### 1.0 INTRODUCTION

To accommodate projected increases in the demand for water, the Texas Water Development Board (TWDB) adopted the Texas Water Plan in 1997. The Texas Water Plan identifies Allens Creek as a potential reservoir site to supply water for the growing populations of Fort Bend and Brazoria counties and central Texas. Water from the lower Brazos River will be diverted to the proposed 142,982 acre-feet reservoir. This project was designed to provide information concerning Brazos River fish communities. To assist in modeling reduced instream flows $15^{\text {th }}, 30^{\text {th }}$, and $50^{\text {th }}$ percentile discharges of the summer and winter seasons were targeted for fish collections. Previous studies documenting fishes occurring near our study reach can be found in Linam et al. (1994) and Winemiller et al. (2000). Studies reporting fish communities of tidal portions and upper reaches of the Brazos River can be found in Johnson (1977), Wilde and Ostrand (1999), Winemiller and Gelwick (1999), and Ostrand and Wilde (2002). McEachran and Fechhelm (1998) lists documented species occurrences in the Brazos River watershed.

This report provides information on habitat characteristics and fish assemblages across $15^{\text {th }}, 30^{\text {th }}$, and $50^{\text {th }}$ percentile discharges in summer and winter. The objectives of this project were to: (1) delineate and photodocument riffle, run, and pool mesohabitats within our study reach; (2) characterize and quantify the fishes occurring in identified mesohabitats; (3) determine indicator species of mesohabitats based on fish distributions; and (4) calculate an Index of Biotic Integrity for the reach.

### 2.0 STUDY AREA

### 2.1 Allens Creek

Allens Creek is a third-order intermittent tributary of the lower Brazos River in southern Austin County, Texas. From its headwaters in Sealy, Allens Creek flows south-southeast and enters the Brazos River 10 km downstream. Year round water flow to the lower portions of Allens Creek is maintained by effluent discharge from the City of Wallis wastewater treatment facility. The proposed reservoir site is located immediately upstream of the FM 1458 road crossing, approximately 900 m above the Allens Creek confluence with the Brazos River.

### 2.2 Brazos River

The headwaters of the Brazos River originate in New Mexico. The river meander eastward across Texas then southeast into the Gulf of Mexico. Several flood control dams and water supply reservoirs are located along the upper reaches of the watershed partially regulating the natural discharge regime. Situated between Austin and Fort Bend counties ( $29^{\circ} 40^{\prime} \mathrm{N}$ and $96^{\circ} 01^{\prime} \mathrm{W}$ ) our study reach is located in the Western Gulf Coastal Plain physiographic region and drains approximately $72,000 \mathrm{~km}^{2}$. Characteristic of its sinuous pattern (Sinuosity Index of 2.16), lateral point bars and deep-water pools dominate the shoreline of our study reach. Rangeland and crop production dominates the land use of the lower Brazos River watershed. A gallery forest dominated by black willow (Salix nigra), sugarberry (Celtis laevigata), elm (Ulmus sp.)
and pecan (Carya sp.) extends along both banks for most of the reach. The study area is described in further detail by McKone et al. 1996.

### 3.0 MATERIALS AND METHODS

### 3.1 Study Reach Delineation

On June 26, 2001 a 10 km study reach was identified during a site visit by representatives from the Texas Water Development Board, Texas Parks and Wildlife, Texas Commission of Environmental Quality (formerly Texas Natural Resources Conservation Commission), U.S. Army Corps of Engineers-Dallas/Fort Worth District, and Texas A\&M University. The study reach was selected as representative habitats in the lower Brazos River downstream of the proposed Allens Creek reservoir. During baseflow conditions on July 11, 2001, representatives from Texas A\&M University and the Texas Commission of Environmental Quality identified sampling sites based on the presence of riffle, run and pool mesohabitats. These mesohabitat-sites were characterized by current velocity, water depth, planform river morphology and the dominant particle size of substrate.

### 3.2 Sampling Schedule

Six collections were completed over a range of river discharges. Collections targeted the $15^{\text {th }}, 30^{\text {th }}$, and $50^{\text {th }}$ percentile discharge of the summer (April through October) and winter (November through March) seasons from September 2001 through August 2002. Target discharges were calculated by the Texas Water Development Board from 60 years of record compiled through the USGS Brazos River at the Richmond, Texas gaging station (\#08114000). Sampling dates and actual discharges during collections are reported in Table 1.

Table 1. Dates and daily discharge of collection periods calculated from USGS Brazos River at Richmond, Texas gaging station (\#08114000).

| Season | Collection Dates | Target Discharge <br> $(\mathrm{cfs})$ | Actual Discharge <br> $($ avg.) |
| :---: | :---: | :---: | :---: |
| Summer $50^{\text {th }}$ | $20-23$ Sept 2001 | 2,630 | 4,043 |
| Summer $30^{\text {th }}$ | $27-30$ Aug 2002 | 1,410 | 1,477 |
| Summer $15^{\text {th }}$ | $13-16$ May 2002 | 924 | 886 |
| Winter $50^{\text {th }}$ | 29 Mar - 01 Apr 2002 | 3,460 | 4,185 |
| Winter $30^{\text {th }}$ | $02-05$ Feb 2002 | 1,710 | 2,623 |
| ${\text { Winter } 155^{\text {th }}}^{08-11 \text { Mar 2002 }}$ | 1,000 | 2,228 |  |

### 3.3 Fish Collections

Seines and gillnets were the primary effective methods used to capture fishes. Nearshore shallow-water areas of each mesohabitat-site were sampled with a $5 \times 1.25 \mathrm{x}$ 1.25 m bag seine of 5 mm bar mesh. Midpoint along each mesohabitat, seines were hauled along at least three contiguous 15 m longitudinal transects until no additional species were captured in two consecutive hauls. The total number of seine hauls was recorded to standardize abundance per $\mathrm{m}^{2}$. Experimental monofilament gillnets measuring 38.1 m long by 1.8 m deep and consisting of five equal sized panels ( $2.5,3.8$, $5.1,6.3$ and 7.6 cm mesh) were used to collect fishes in deep-water habitats. Three to five gillnets were set overnight for a total of $9-15$ sets per collection period. Gillnets were set with one end anchored into a riverbank or large woody debris and set at a $45^{\circ}$ or $315^{\circ}$ angle with the shoreline. Backwaters support the vast proportion of fishes in large rivers (Stalnaker et al. 1989), so gillnets were typically set to target backwater areas within mesohabitat-sites. Gillnet captures were standardized as abundance per $\mathrm{m}^{2}$ of net.

Deep-water areas, large aggregations of woody debris, and mesohabitat-sites dominated by large woody debris were sampled with a boat-mounted electrofisher. We used a Coffelt model VVP-2C electrofisher powered by a 5000 watt Honda generator mounted onto a $4.3-\mathrm{m}$ aluminum jon boat powered by a 15 -horsepower Mercury outboard. Fishes were captured only in areas of large aggregations of woody debris and mesohabitat-sites dominated by large woody debris during the winter $30^{\text {th }}$ and summer $15^{\text {th }}$ percentile discharge collections. Due to technical difficulties with electrofishing equipment, samples were not collected in the woody debris field near the downstream end of our study reach (mesohabitat-site $H$ ) during the winter $30^{\text {th }}$ percentile collections. Electrofishing catch was standardized as abundance per $\mathrm{m}^{2}$ sampled.

Three baited funnel-type minnow traps of 7.62 mm mesh and 2.54 cm funnel openings were also used to collect fishes during the winter $50^{\text {th }}, 30^{\text {th }}, 15^{\text {th }}$ and summer $15^{\text {th }}$ percentile discharge rates. Minnow traps were deployed in large aggregations of woody debris across the study reach and allowed to fish for approximately 72 hours. Additionally, during the summer $15^{\text {th }}$ and $30^{\text {th }}$ percentile discharge collections, two 61 cm diameter hoopnets of 2.54 cm mesh and two 91.44 cm diameter hoopnets of 2.54 cm
mesh were set. Hoopnets were baited with a can of catfood, positioned with the openings facing downstream and allowed to fish for 72 hours. Hoopnet and minnow trap captures were standardized as abundance per $\mathrm{m}^{2}$ sampled by their openings.

Captured individuals that were rare, threatened, or endangered and large common fishes were identified and immediately returned to the river. All other fishes were euthanized in tricane (MS-222), fixed in 10\% formalin, and returned to the lab for enumeration. With the exception of bowfin (Amia calva) and spotted gar (Lepidosteus oculatus), several individuals of each species captured was catalogued as voucher specimen into the Texas Cooperative Wildlife Collections located on the campus of Texas A\&M University.

### 3.4 Habitat Assessments

Physicochemical parameters were measured immediately following fish collections. Temperature ( ${ }^{\circ} \mathrm{C}$ ), conductivity ( $\mu \mathrm{S} / \mathrm{cm}$ ), dissolved oxygen concentration $(\mathrm{mg} / \mathrm{L})$ and saturation (\%) were measured in the center of each sampling area with a YSI-85 (Yellow Springs Instrument) multimeter. Water depth and velocity were measured at 3 equidistant points along a diagonal bisecting each area seined or electrofished. Single values for water depth and current velocity of gillnet, hoopnet or minnow trap sites were measured in the center of the sampled area. Water depths less than 150 cm were measured using a graduated wading rod. Depths greater than 150 cm were measured using a Speedtech ${ }^{\circledR}$ sonar depth meter. Flow was measured at 0.6 times the water depth using a Marsh-McBirney Flowmate 2000 electromagnetic flow meter. At large woody debris habitats, flows were measured several feet upstream of the structure. Areas sampled were photodocumented during the winter $30^{\text {th }}$ or summer $30^{\text {th }}$ percentile discharge collections.

### 3.5 Index of Biotic Integrity (IBI)

The Index of Biotic Integrity (IBI) assesses attributes of the fish assemblage to determine water quality and condition of aquatic ecosystems (Karr 1981). We calculated an IBI for our study reach using metrics developed by Winemiller and Gelwick (1999) for the Brazos-Navasota River watershed (Tables 2-5). Since reference data for large, undisturbed rivers in Texas were unavailable (Bayer et al. 1992), we compared our IBI scores to scores calculated for sites sampled in autumn along the mainstem of the lower Brazos River by Winemiller and Gelwick (1999). We calculated four scores of our study reach: (1) seine captures during autumn collections; (2) captures in all gears during autumn collections; (3) seine captures across the six rates of discharge; and (4) captures in all gears across the six rates of discharge.

Table 2. Outline of the IBI metrics and scoring criteria as adapted for the Brazos River in central Texas (from Winemiller and Gelwick 1999).

|  | Scoring Criteria |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 10 | 7 | 5 | 2 | 0 |
| Species Richness and Composition Metrics: |  |  |  |  |  |
| \# native species | $21+$ | 16-20 | 10-15 | 5-9 | 0-4 |
| \# darter species | 4+ | 3 | 2 | 1 | 0 |
| \# sunfish species | 4+ | 3 | 2 | 1 | 0 |
| \# sucker species | $1+$ | - | - | - | 0 |
| \# intolerant species | 8+ | 6-7 | 3-5 | 1-2 | 0 |
| \% tolerant species | 0-49 | 50-79 | 80-89 | 90-94 | 95-100 |
| \% mosquitofish | 0-1 | 2-9 | 10-19 | 20-29 | 30-100 |
| Trophic Function Metrics: |  |  |  |  |  |
| \% omnivores | 0-75 | 76-79 | 80-89 | 90-94 | 95-100 |
| \% invertivores | 25-100 | 20-24 | 11-19 | 6-10 | 0-5 |
| \% carnivores | 7-100 | 4-6 | 2-3 | 0.1-1 | 0 |

Table 3. Assignment of fish species for the species richness and composition metrics of the IBI (adopted from Winemiller and Gelwick 1999).

| Non-native species | Cyprinus carpio |
| :--- | :--- |
| Darters | Etheostoma gracile, Noturus gyrinus |
| Suckers | Carpoides carpio, Ictiobus bubalus |
| Sunfish | Lepomis cynaellus, L. gulosus, L. humilis, L. macrochirus, L. <br> marginatus, L. megalotis, L. microlophus, L. punctatus, <br> Pomoxis annularis |
| Intolerant species | Cyprinus carpio, Etheostoma gracile, Labidesthes sicculus, <br> Lepomis megalotis, Lythrurus fumeus, Menidia beryllina, <br> Notropis buchanani, Notropis shumardi, Noturus gyrinus, <br> Opsopeoedus emiliae |
|  | Amia calva, Aplodinotus grunniens, Cyprinella lutrensis, <br> Corant species |
|  | Ictalurus punctatus, Lepisosteus oculatus, L. osseas, Lepomis <br> cyanellus, L. gulosus, L. macrochirus, Pimephales vigilax |
|  |  |

Table 4. Assignment of fish species from trophic structure metrics of IBI (adopted from Winemiller and Gelwick 1999).

| Omnivores | Carpoides carpio, Cyprinus carpio, Cyprinella lutrensis, <br> Dorosoma cepedianum, Mugil cephalus, Pimephales vigilax |
| :--- | :--- |
| Invertivores | Aphredodreus sayanus, Aplodinotus grunniens, Cyprinella <br> venusta, Dorosoma petenense, Etheostoma gracile, Ictiobus <br> bubalus, Macrhybopsis aestivalis, M. storeriana, Fundulus |
|  | notatus, Gambusia affinis, Ictiobus bubalus, Labidesthes <br> sicculus, Lepomis cyanellus, L. humilis, L. macrochirus, L. <br>  <br> marginatus, L. megalotis, L. microlophus, L. punctatus, <br>  <br> Lythrurus fumeus, Menidia beryllina, Notropis buchanani, N. <br> oxyrhynchus, N. shumardi, Noturus gyrinus, Opsopeoedus <br> emiliae |
| Top carnivores | Amia calva, Ictalurus furcatus, Ictalurus punctatus, <br>  <br> Lepisosteus osseus, L. oculatus, Lepomis gulosus, <br> Micropterus puntulatus, Micropterus salmoides, Pomoxis <br> annularis, Pyliodictus olivaris |

Table 5. Interpretation of IBI scores (from Winemiller and Gelwick 1999).

| IBI Score | Assessment | Fish Community and Stream Attributes |
| :---: | :--- | :--- |
| 65-100 | Excellent | Comparable to the best situations with minimal human <br> disturbance; most of the regionally expected species for <br> habitat and stream size, including the most intolerant <br> forms, are present with a balanced trophic structure. |
| $50-64$ | Good | Species richness somewhat below expectation, especially <br> due to the loss of the most intolerant forms; some species, <br> especially top carnivores, are present with less than <br> optimal abundances; trophic structure may show signs of <br> imbalance. |
| $30-49$ | Fair | Signs of additional deterioration include decreased species <br> richness, loss of intolerant forms, increased abundance of <br> tolerant species, and/or highly skewed trophic structure <br> (e.g., greater frequency of omnivores and lower frequency <br> of invertebrate feeders and carnivores. |
| $20-29$ | Poor | Relatively few species; dominated by omnivores, tolerant <br> forms, and habitat generalists; few or no top carnivores. |
| $0-19$ | Very Poor | Very few species present, mostly exotics or tolerant <br> forms; few large or old fish; diseased fish may be <br> common. |

### 3.6 Indicator Species Analysis

We performed an indicator species analysis (Dufrêne and Legendre 1997) based on percent abundances in collections and percent occurrence among collections to test the probability that species were indicators of pool, run, riffle, and tributary confluence mesohabitats. We calculated species abundance per $\mathrm{m}^{2}$ sampled in each mesohabitattype for each of our six collection periods. Two separate analyses were performed with PC-ORD (McCune and Mefford 1997): (1) using only those species exceeding $1 \%$ of total collections; and (2) including all species regardless of abundance.

### 4.0 RESULTS AND CONCLUSIONS

### 4.1 Mesohabitat-Site Delineation

Eleven sites were identified based upon mesohabitat delineations. Five runs, 4 pools, 1 riffle and a tributary confluence were each designated by a unique mesohabitatsite code (Figure 1). The presence of pool, run, or riffle mesohabitats did not vary across our six collection discharges ( $886-4185 \mathrm{cfs}$ ). However, slight reductions in mesohabitat volume (water surface area and depth) were observed with decreasing discharge. The lower reaches of Allens Creek was hydrologically connected to waters of the Brazos River during collections at all targeted discharges. However, during our summer $15^{\text {th }}$ percentile collections, fish movement between the Brazos River and Allens Creek was likely impeded by the combined effects of a low river stage and high sediment aggradation which acted as a low-water dam across the mouth of Allens Creek. Additionally, the large woody debris aggregation at the FM 1093 bridge crossing was elevated above the water on a sediment bar and did not provide woody habitat for fish during the summer $15^{\text {th }}$ percentile discharge.


Figure 1. Sketch map of Brazos River study reach with mesohabitat-sites indicated by a letter code and sampling locations by a numeric code.

### 4.2 Physicochemical Parameters

Mean daily discharge ranged from 1,792 to $17,300 \mathrm{cfs}$ (from 82 years of record), compared to a range of 886 to $4,185 \mathrm{cfs}$ during our collection periods (Figure 2). Averaged across all sites, water temperature ranged from 13.8 to $31.4^{\circ} \mathrm{C}$, conductivity ranged 467.5 to $1059.0 \mu \mathrm{~S} / \mathrm{cm}$, dissolved oxygen concentration from 6.72 to 13.67 and saturation from 76.2 to $117.5 \%$ for each collection period (Table 6).

Water depths and current velocities of each sampling location are reported for each collection period in Tables 7 and 8. Mean depth and current velocity measurements of mesohabitat within each collection period are reported in Table 9. Because gillnets were generally deployed in deep backwaters and not areas representative of their respective mesohabitat-site, we did not include gillnet depths and velocities in our overall calculations of the mean. Mean current velocities were related to mesohabitat types. Pool mesohabitat-sites were generally characterized by minimal velocities (mean 14.2; range 7.7 to $20.7 \mathrm{~cm} / \mathrm{s}$ ). Runs were characterized by moderate velocities (21.3; 15.4 to $27.9 \mathrm{~cm} / \mathrm{s}$ ) and riffles by the highest velocities ( $34.1 ; 20.0$ to $66.0 \mathrm{~cm} / \mathrm{s}$ ). Velocities of the Allens Creek confluence site were negligible due to a backwater effect by riverflow of the Brazos River. Mean water depths of areas seined were 38.6, 50.6, 50.8 , and 38.6 cm in pool, run, riffle and tributary confluence mesohabitats, respectively.

### 4.3 Fish Species and Mesohabitat Use

A total of 44,122 individuals representing 43 species from 14 families were collected across our 6 collection periods (Table 10). Red shiners (Cyprinella lutrensis) and bullhead minnows (Pimephales vigilax) accounted for $67.4 \%$ and $16.9 \%$ of our collections, respectively. Other common species (abundances exceeding $1 \%$ of overall collections) were ghost shiner (Notropis buchanani), silverband shiner (N. shumardi), striped mullet (Mugil cephalus), and mosquitofish (Gambusia affinis). Three individuals of sharpnose shiner (Notropis oxyrhynchus) were collected in the confluence of Allens Creek (mesohabitat-site AC) during our summer $50^{\text {th }}$ percentile discharge collections. The sharpnose shiner was recently proposed as a candidate species for federal listing by the U.S. Fish and Wildlife Service (2002). Bubble graphs of fish species collections per sampled location for each of the targeted discharge rates are provided in Figures 3-8. Species and sampling location codes used in the bubble graphs are listed in Table 11. Photos of representative habitats sampled are provided in Figures 9-40. All photos are looking upriver. A list of species documented to occur in the Brazos River near our study reach is provided in Table 12 (Linam et. al 1994, Winemiller et. al. 2000).


Figure 2. Historical (based on 82 years of record) and mean daily discharge recorded during the study period (September 01, 2001-August 31, 2002) at the USGS Brazos River at Richmond, Texas gage (station \#08114000).

Table 6. Physicochemical parameters for each collection period (reported as the mean of all sampling locations).

| Season | Sampling Dates | Mean Daily <br> Discharge $(\mathrm{cfs})$ | Temperature <br> $\left({ }^{\circ} \mathrm{C}\right)$ | Conductivity <br> $(\mu \mathrm{S} / \mathrm{cm})$ | Dissolved Oxygen <br> Concentration <br> $(\mathrm{mg} / \mathrm{L})$ | Saturation (\%) |
| :--- | :--- | :--- | :---: | :---: | :---: | :---: |
| Summer 50 th | September 20-23, 2001 | 4043 | 28.9 | 492.2 | 8.79 | 76.2 |
| Summer 30 $0^{\text {th }}$ | August 27-30, 2002 | 1477 | 31.4 | 1059.0 | 6.72 | 91.6 |
| Summer $15^{\text {th }}$ | May 13-16, 2002 | 886 | 26.2 | 856.3 | 8.20 | 107.7 |
| Winter 50 $0^{\text {th }}$ | March 29-April 1, 2002 | 4185 | 20.9 | 467.5 | 8.18 | 91.8 |
| Winter $30^{\text {th }}$ | February 2-5, 2002 | 2533 | 13.8 | 589.4 | 13.67 | 110.2 |
| Winter $15^{\text {th }}$ | March 8-11, 2002 | 2228 | 17.5 | 569.4 | 11.23 | 117.5 |

Table 7. Depth and current velocities of sampling locations during summer collection periods. Habitat codes correspond with bubble graph and are described in Table 11.

| Habitat-Code | $50^{\text {th }}$ Percentile |  | $30^{\text {th }}$ Percentile |  | $15^{\text {th }}$ Percentile |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Depth (cm) | Velocity (cm/s) | Depth <br> (cm) | Velocity (cm/s) | Depth (cm) | Velocity (cm/s) |
| 1 | 165.4 | 19.5 |  |  |  |  |
| 2 | 53.7 | 45.3 | 51.3 | 29.3 | 40.7 | 23.3 |
| 3 |  |  |  |  |  |  |
| 4 | 280.0 | 4.0 | 200.0 | 55.0 | 93.0 | 21.0 |
| 5 | 67.7 | 9.0 | 73.7 | 17.3 | 77.0 | 16.3 |
| 6 | 230.0 | 30 | 120.0 | 49.0 | 98.0 | 37.0 |
| 7 | 161.3 | 26.3 | 250.0 | 10.0 |  |  |
| 8 | 33.3 | 31.7 | 93.0 | 66.0 | 39.0 | 23.7 |
| 9 |  |  |  |  | 25.3 | 0.0 |
| 10 |  |  | 150.0 | 45.0 |  |  |
| 11 |  |  |  |  |  |  |
| 12 | 17.7 | 7.7 | 40.3 | 33.7 | 15.7 | 19.3 |
| 13 |  |  | 335.0 | 9.0 |  |  |
| 14 |  |  | 120.0 | 52.8 | 101.0 | 47.0 |
| 15 | 15.0 | -0.3 | 91.2 | 16.5 | 22.0 | 15.7 |
| 16 | 88.3 | -0.7 | 62.3 | 3.0 | 38.7 | 1.0 |
| 17 | 400.0 | -34.0 | 78.3 | 14.0 | 84.3 | 13.7 |
| 18 | 190.0 | 8.0 |  |  |  |  |
| 19 |  |  | 221.5 | 0.0 | 128.2 | 2.1 |
| 20 (LWD) | 156 | 62.0 | 230.0 | 0.0 | 208.4 | -1.5 |
| 20 |  |  | 210.0 | 45.0 | 130.0 | 71.0 |
| 21 |  |  |  |  | 123.0 | 33.0 |
| 22 | 44.2 | 12.7 | 23.3 | 13.3 | 34.0 | 12.3 |
| 23 |  |  |  |  | 123.7 | 46.3 |
| 24 |  |  | 140.0 | 89.0 | 125.0 | 58.0 |
| 25 | 71.7 | 29.7 | 18.0 | 22.7 | 34.7 | 16.7 |
| 26 | 33.7 | 16.3 | 48.0 | 20.0 | 48.7 | 14.7 |
| 27 | 270.0 | -6.0 | 441.0 | -5.5 | 160.0 | -4.0 |
| 28 | 310.0 | 30.0 | 116.9 | 25.5 | 56.2 | 10.9 |
| 29 | 38.0 | 25.7 | 65.7 | -5.0 | 53.7 | -1.7 |
| 30 | 64.7 | 19.1 | 105.2 | 13.1 | 40 | 8.0 |

Table 8. Depth and current velocities of sampling locations during winter collection periods. Habitat codes correspond with bubble graph and are described in Table 11.

| Habitat-Site | $50^{\text {th }}$ Percentile |  | $30^{\text {th }}$ Percentile |  | $15^{\text {th }}$ Percentile |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Depth <br> (cm) | Velocity (cm/s) | Depth <br> (cm) | Velocity (cm/s) | Depth <br> (cm) | Velocity (cm/s) |
| 1 |  |  | 153.2 | 39.0 |  |  |
| 2 | 45.7 | 22.3 | 43.7 | 23.7 | 35.7 | 20.3 |
| 3 |  |  | 216.7 | 52.3 |  |  |
| 4 |  |  |  |  | 230.7 | 31.0 |
| 5 | 44.7 | 10.7 | 40.3 | 8.0 | 27.3 | 7.7 |
| 6 | 190.0 | 49.0 | 210.0 | 27.0 | 131.0 | 60.0 |
| 7 | 35.0 | 21.0 | 185.3 | 30.0 | 145.0 | -17.0 |
| 8 | 64.0 | 35.0 | 36.3 | 28.3 | 39.3 | 20.0 |
| 9 |  |  | 22.7 | 31.7 | 32.3 | 30.7 |
| 10 |  |  | 142.7 | 40.3 |  |  |
| 11 |  |  | 150.0 | 35.0 |  |  |
| 12 | 15.0 | 15.7 | 53.3 | 43.7 | 28.3 | 22.0 |
| 13 | 202.0 | 10.5 | 146.7 | 50.7 | 122.0 | 22.0 |
| 14 |  |  |  |  |  |  |
| 15 | 31.0 | 15.0 | 13.7 | 2.0 | 34.0 | 39.7 |
| 16 | 45.5 | 21.0 | 58.2 | -0.3 | 32.2 | 20.2 |
| 17 | 73.0 | 8.7 | 101.0 | -1.0 | 56.7 | 16.7 |
| 18 |  |  |  |  |  |  |
| 19 | 355.0 | -7.0 | 290.0 | -5.5 | 255.0 | -9.5 |
| 20 (LWD) | 300.0 | -2.0 | 105.0 | -1.0 | 190.0 | 0.0 |
| 20 |  |  |  |  | 190.0 | 0.0 |
| 21 |  |  | 122.0 | 56.0 |  |  |
| 22 | 26.7 | 14.5 | 47.5 | 17.5 | 26.0 | 8.7 |
| 23 |  |  |  |  |  |  |
| 24 | 80.0 | 42.0 | 121.0 | 26.0 | 62.0 | 33.0 |
| 25 | 186.2 | 26.5 | 45.0 | 22.3 | 58.0 | 13.7 |
| 26 | 23.0 | 16.0 | 51.3 | 11.0 | 39.0 | 5.7 |
| 27 | 180.0 | -4.0 | 159.0 | -1.0 | 200.0 | 1.0 |
| 28 | 66.0 | 16.2 | 81.7 | 14.0 | 113.0 | 26.4 |
| 29 | 97.9 | 23.5 | 74.5 | 0.0 | 54.4 | -3.0 |
| 30 | 106.0 | 8.9 | 39.2 | 14.7 | 70.4 | 1.9 |

Table 9. Mean water depth and current velocity measurements of mesohabitats during collections. (Note: because gillnets targeted backwaters and not areas representative of mesohbabitats, depth and velocity measurements of gillnetted areas were not included in these calculations.)

| Season | Water Depth (cm) |  |  |  | Current Velocity (cm/s) |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Pool | Run | Riffle | Tributary Confluence | Pool | Run | Riffle | Tributary Confluence |
| Summer $50^{\text {th }}$ | 42.5 | 45.9 | 33.3 | 64.2 | 16.7 | 27.9 | 31.7 | 14.8 |
| Summer $30^{\text {th }}$ | 44.3 | 53.9 | 93.0 | 105.2 | 20.7 | 23.7 | 66.0 | 13.2 |
| Summer $15^{\text {th }}$ | 45.4 | 44.7 | 39.0 | 40.0 | 14.8 | 18.5 | 23.7 | 4.0 |
| Winter $50^{\text {th }}$ | 30.4 | 63.9 | 64.0 | 106.0 | 14.1 | 19.2 | 35.0 | 7.5 |
| Winter $30{ }^{\text {th }}$ | 38.7 | 49.1 | 36.3 | 64.2 | 7.7 | 23.2 | 28.3 | 14.7 |
| Winter $15^{\text {th }}$ | 30.2 | 45.9 | 39.3 | 70.4 | 11.0 | 15.4 | 20.0 | 1.9 |

Table 10. Total species abundance across collection periods.

| Species | Species Abundance/Collection Period |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Summer |  |  | Winter |  |  | Total |
|  | $15^{\text {th }}$ | $30^{\text {th }}$ | $50^{\text {th }}$ | $15^{\text {th }}$ | $30^{\text {th }}$ | $50^{\text {th }}$ |  |
| Amiidae |  |  |  |  |  |  |  |
| Amia calva (bowfin) | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| Aphredoderidae |  |  |  |  |  |  |  |
| Aphredodreus sayanus (pirate perch) | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| Atherinidae |  |  |  |  |  |  |  |
| Labidesthes sicculus (brook silverside) | 0 | 0 | 4 | 0 | 0 | 0 | 4 |
| Menidia beryllina (inland silverside) | 6 | 3 | 22 | 1 | 0 | 1 | 33 |
| Catostomidae |  |  |  |  |  |  |  |
| Carpoides carpio (river carpsucker) | 8 | 6 | 16 | 5 | 3 | 2 | 40 |
| Ictiobus bubalus (smallmouth buffalo) | 3 | 6 | 0 | 3 | 3 | 1 | 16 |
| Centrarchidae |  |  |  |  |  |  |  |
| Lepomis cyanellus (green sunfish) | 1 | 0 | 0 | 7 | 5 | 0 | 13 |
| Lepomis gulosus (warmouth) | 0 | 0 | 1 | 0 | 2 | 0 | 3 |
| Lepomis humilis (orangespotted sunfish) | 0 | 5 | 4 | 2 | 2 | 0 | 13 |
| Lepomis macrochirus (bluegill sunfish) | 2 | 2 | 2 | 6 | 1 | 1 | 14 |
| Lepomis megalotis (longear sunfish) | 2 | 2 | 3 | 1 | 11 | 3 | 22 |
| Lepomis microlophus (redear sunfish) | 0 | 4 | 4 | 0 | 0 | 0 | 8 |
| Lepomis punctatus (spotted sunfish) | 0 | 0 | 0 | 0 | 2 | 0 | 2 |
| Lepomis hybrid (hybrid sunfish) | 0 | 0 | 0 | 1 | 1 | 0 | 2 |
| Lepomis sp. (juvenile sunfish) | 0 | 3 | 11 | 0 | 0 | 1 | 15 |
| Micropterus punctulatus (spotted bass) | 1 | 0 | 0 | 0 | 0 | 2 | 3 |
| Micropterus salmoides (largemouth bass) | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Pomoxis annularis (white crappies) | 0 | 1 | 4 | 0 | 0 | 0 | 5 |
| Clupeidae |  |  |  |  |  |  |  |
| Alosa crysochloris (skipjack herring) | 0 | 3 | 0 | 0 | 10 | 0 | 3 |
| Dorosoma cepedianum (gizzard shad) | 2 | 2 | 41 | 7 | 10 | 12 | 74 |
| Dorosoma petenense (threadfin shad) | 2 | 70 | 60 | 3 | 10 | 3 | 148 |

Table 10. Species abundance across collection periods (continued).

| Cyprinidae |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Cyprinella lutrensis (red shiner) | 5006 | 1611 | 2558 | 9664 | 4712 | 6172 | 29723 |
| Cyprinella venusta <br> (blacktail shiner) | 2 | 2 | 0 | 1 | 0 | 1 | 6 |
| Cyprinus carpio (common carp) | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Lythrurus fumeus (ribbon shiner) | 0 | 0 | 2 | 0 | 1 | 0 | 3 |
| Macrhybopsis aestivalis (speckled chub) | 0 | 10 | 52 | 11 | 27 | 45 | 145 |
| Machrybopsis storeriana (silver chub) | 2 | 1 | 39 | 0 | 3 | 0 | 45 |
| Notropis buchanani (ghost shiner) | 62 | 0 | 75 | 316 | 64 | 446 | 963 |
| Notropis oxyrhynchus (sharpnose shiner) | 0 | 0 | 3 | 0 | 0 | 0 | 3 |
| Notropis shumardi (silverband shiner) | 134 | 11 | 311 | 659 | 83 | 934 | 2132 |
| Opsopoeodus emiliae <br> (pugnose minnow) | 1 | 0 | 1 | 0 | 0 | 0 | 2 |
| Pimephales vigilax (bullhead minnow) | 156 | 266 | 867 | 1660 | 1039 | 3448 | 7436 |
| Fundulidae |  |  |  |  |  |  |  |
| Fundulus notatus <br> (blackstripe topminnow) | 1 | 0 | 0 | 1 | 0 | 0 | 2 |
| Ictaluridae |  |  |  |  |  |  |  |
| Ictalurus furcatus <br> (blue catfish) | 3 | 6 | 8 | 4 | 6 | 1 | 28 |
| Ictalurus punctatus (channel catfish) | 7 | 2 | 17 | 4 | 20 | 12 | 62 |
| Noturus gyrinus (tadpole madtom) | 0 | 0 | 0 | 0 | 1 | 1 | 2 |
| Pylodictis olivaris <br> (flathead catfish) | 3 | 1 | 2 | 0 | 7 | 1 | 14 |
| Lepisosteidae |  |  |  |  |  |  |  |
| Lepidosteus oculatus (spotted gar) | 4 | 8 | 3 | 11 | 4 | 29 | 59 |
| Lepidosteus osseus (longnose gar) | 111 | 9 | 4 | 34 | 8 | 42 | 208 |
| Mugilidae |  |  |  |  |  |  |  |
| Mugil cephalus (striped mullet) | 15 | 0 | 0 | 2 | 25 | 1079 | 1121 |
| Percidae |  |  |  |  |  |  |  |
| Etheostoma gracile <br> (slough darter) | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| Poeciliidae |  |  |  |  |  |  |  |
| Gambusia affinis (mosquitofish) | 833 | 181 | 465 | 41 | 147 | 74 | 1741 |
| Sciaenidae |  |  |  |  |  |  |  |
| Aplodinotus grunniens (freshwater drum) | 2 | 2 | 0 | 0 | 0 | 1 | 5 |
| Totals | 6369 | 2219 | 4580 | 12445 | 6197 | 12312 | 44122 |

Table 11. Species and sampling location codes reported in bubble graphs.

| Code | Species | Habitat |
| :---: | :---: | :---: |
| 1 | Amia calva (bowfin) | A - run; all |
| 2 | Aphredodreus sayanus (pirate perch) | A - run; left margin |
| 3 | Labidesthes sicculus (brook silverside) | A - run; mid channel |
| 4 | Menidia beryllina (inland silverside) | A - run; right margin |
| 5 | Carpoides carpio (river carpsucker) | $B-$ pool; left margin |
| 6 | Ictiobus bubalus (smallmouth buffalo) | B - pool; mid channel |
| 7 | Lepomis cyanellus (green sunfish) | $B$ - pool; right margin |
| 8 | Lepomis gulosus (warmouth) | C-riffle; mid channel |
| 9 | Lepomis humilis (orangespotted sunfish) | C-riffle; right margin |
| 10 | Lepomis macrochirus (bluegill sunfish) | D - run; left margin |
| 11 | Lepomis megalotis (longear sunfish) | D - run; mid channel |
| 12 | Lepomis microlophus (redear sunfish) | D - run; right margin |
| 13 | Lepomis punctatus (spotted sunfish) | E-pool; left margin |
| 14 | Lepomis hybrid (hybrid sunfish) | E-pool; mid channel |
| 15 | Lepomis sp. (juvenile sunfish TL<20mm) | E-pool; right margin |
| 16 | Micropterus punctulatus (spotted bass) | F - backwater; left bank |
| 17 | Micropterus salmoides (largemouth bass) | $F$ - run; left margin |
| 18 | Pomoxis annularis (white crappie) | F - run; mid channel |
| 19 | Alosa crysochloris (skipjack herring) | F - run; right margin |
| 20 | Dorosoma cepedianum (gizzard shad) | G - pool; left margin/LWD |
| 21 | Dorosoma petenense (threadfin shad) | G - pool; mid channel |

Table 11. Species and sampling location codes reported in bubble graphs (continued).
22 Cyprinella lutrensis (red shiner) G - pool; right channel

Cyprinella venusta (blacktail shiner)
Cyprinus carpio (common carp)
Lythrurus fumeus (ribbon shiner)
Macrhybopsis aestivalis (speckled chub)
Machrybopsis storeriana (silver chub)
Notropis buchanani (ghost shiner)
Notropis oxyrhynchus (sharpnose shiner)
Notropis shumardi (silverband shiner)
Opsopoeodus emiliae (pugnose minnow)
Pimephales vigilax (bullhead minnow)
Fundulus notatus (blackstripe topminnow)
Ictalurus furcatus (blue catfish)
Ictalurus punctatus (channel catfish)
Noturus gyrinus (tadpole madtom)
Pylodictis olivaris (flathead catfish)
Lepidosteus oculatus (spotted gar)
Lepidosteus osseus (longnose gar)
Mugil cephalus (striped mullet)
Etheostoma gracile (slough darter)
Gambusia affinis (mosquitofish)
Aplodinotus grunniens (freshwater drum)


Figure 3. Total number of fishes collected in the Brazos River during summer $50^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 4. Total number of fishes collected in the Brazos River during summer $30^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 5. Total number of fishes collected in the Brazos River during summer $15^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 6. Total number of fishes collected in the Brazos River during winter $50^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 7. Total number of fishes collected in the Brazos River during winter $30^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 8. Total number of fishes collected in the Brazos River during winter $15^{\text {th }}$ percentile discharge collections. Number of fishes is indicated at the intersections of species and sampling location codes and also by the relative size of bubbles centered at intersections. Zeros indicate the species was not collected from the habitat. No number at an intersection indicates the habitat was not sampled. Species and sampling location codes are in Table 9.


Figure 9. Mesohabitat A - left margin (sampling location code - 2; all photos looking upriver); Aug 27-30, 2002; Depth $(D)=51.3 \mathrm{~cm}$, Velocity $(V)=29.3 \mathrm{~cm}$.


Figure 10. Mesohabitat A - midchannel (code - 1);
Feb 2-5, 2002; $D=216.7 \mathrm{~cm}, V=52.3 \mathrm{~cm}$.


Figure 11. Mesohabitat A - right margin (code - 4);
Aug 27-30, 2002; $\mathrm{D}=140 \mathrm{~cm}, \mathrm{~V}=30 \mathrm{~cm}$.


Figure 12. Mesohabitat B - left margin (code - 5);
Aug 27-30, 2002; $\mathrm{D}=73.7 \mathrm{~cm}, \mathrm{~V}=17.3 \mathrm{~cm}$.


Figure 13. Mesohabitat B - midchannel (code - 6); Aug 27-30, 2002; $D=120 \mathrm{~cm}, \mathrm{~V}=49 \mathrm{~cm}$.


Figure 14. Mesohabitat B - right margin (code - 7);
Aug 27-30, 2002; $\mathrm{D}=250 \mathrm{~cm}, \mathrm{~V}=10 \mathrm{~cm}$.


Figure 15. Mesohabitat C - midchannel (code - 8);
Aug 27-30, 2002; $\mathrm{D}=93.0 \mathrm{~cm}, \mathrm{~V}=66.0 \mathrm{~cm}$.


Figure 16. Mesohabitat C - right margin (code - 9);
Feb 2-5, 2002; $\mathrm{D}=22.7 \mathrm{~cm}, \mathrm{~V}=31.7 \mathrm{~cm}$.


Figure 17. Mesohabitat D - left margin (code - 10);
Aug 27-30, 2002; $D=150 \mathrm{~cm}, \mathrm{~V}=45 \mathrm{~cm}$.


Figure 18. Mesohabitat D - midchannel (code - 11 in background and code - 14 in foreground); Aug 27-30, 2002; code $-11: \mathrm{D}=\mathrm{n} / \mathrm{a}, \mathrm{V}=\mathrm{n} / \mathrm{a}$; code -14 : $\mathrm{D}=110 \mathrm{~cm}, \mathrm{~V}=70 \mathrm{~cm}$.


Figure 19. Mesohabitat D - right margin (code - 12);
Feb 2-5, 2002; $\mathrm{D}=53.3 \mathrm{~cm}, \mathrm{~V}=43.7 \mathrm{~cm}$.


Figure 20. Mesohabitat E - left margin (code - 13);
Aug 27-30, 2002; $\mathrm{D}=160 \mathrm{~cm}, \mathrm{~V}=15 \mathrm{~cm}$.


Figure 21. Mesohabitat E - midchannel and right margin (codes - 14 and 15); Aug 27-30, 2002; code $-14: \mathrm{D}=110 \mathrm{~cm}, \mathrm{~V}=70 \mathrm{~cm}$; code $-15: \mathrm{D}=32.3 \mathrm{~cm}, \mathrm{~V}=32.0 \mathrm{~cm}$.


Figure 22. Mesohabitat F - left margin (code - 16; backwater);
Aug 27-30, 2002; $\mathrm{D}=62.3 \mathrm{~cm}, \mathrm{~V}=3.0 \mathrm{~cm}$.


Figure 23. Mesohabitat F - left margin (code - 17); Aug 27-30, 2002; $D=n / a, V=n / a$.


Figure 24 Mesohabitat F - LWD (code - 17; in background) and midchannel (code - 18; in foreground); Aug 27-30, 2002; code -17 : $\mathrm{D}=78.3 \mathrm{~cm}, \mathrm{~V}=14.0 \mathrm{~cm}$; code $-18: \mathrm{D}=\mathrm{n} / \mathrm{a}, \mathrm{V}=\mathrm{n} / \mathrm{a}$.


Figure 25. Mesohabitat F - right margin (code - 19; upstream);
Aug 27-30, 2002; $\mathrm{D}=280 \mathrm{~cm}, \mathrm{~V}=-2 \mathrm{~cm}$.


Figure 26. Mesohabitat F - right margin (code - 19; downstream);
Aug 27-30, 2002; $D=163 \mathrm{~cm}, \mathrm{~V}=2 \mathrm{~cm}$.


Figure 27. Mesohabitat F - right margin (code - 19; LWD);
Aug 27-30, 2002; $D=n / a, V=n / a$.


Figure 28. Mesohabitat G-left margin (code - 20);
Aug 27-30, 2002; $\mathrm{D}=210 \mathrm{~cm}, \mathrm{~V}=45 \mathrm{~cm}$.


Figure 29. Mesohabitat G - left margin (code - 20; LWD); Aug 27-30, 2002; $\mathrm{D}=230 \mathrm{~cm}, \mathrm{~V}=0 \mathrm{~cm}$.


Figure 30. Mesohabitat G - midchannel (code - 21);
Aug 27-30, 2002; $\mathrm{D}=110 \mathrm{~cm}, \mathrm{~V}=49 \mathrm{~cm}$.


Figure 31. Mesohabitat G - right margin (code - 22);
Aug 27-30, 2002; $\mathrm{D}=23.3 \mathrm{~cm}, \mathrm{~V}=13.3 \mathrm{~cm}$.


Figure 32. Mesohabitat H - midchannel (code - 23);
Aug 27-30, 2002; $\mathrm{D}=140 \mathrm{~cm}, \mathrm{~V}=89 \mathrm{~cm}$.


Figure 33. Mesohabitat H - right margin (code - 25);
Aug 27-30, 2002; $\mathrm{D}=18.0 \mathrm{~cm}, \mathrm{~V}=22.7 \mathrm{~cm}$.


Figure 34. Mesohabitat I - left margin (code - 26); Aug 27-30, 2002; $\mathrm{D}=48.0 \mathrm{~cm}, \mathrm{~V}=20.0 \mathrm{~cm}$.


Figure 35. Mesohabitat I - right margin (code - 27);
Aug 27-30, 2002; $\mathrm{D}=132 \mathrm{~cm}, \mathrm{~V}=-8 \mathrm{~cm}$.


Figure 36. Mesohabitat J - left margin (code - 28); Aug 27-30, 2002; $\mathrm{D}=81.7 \mathrm{~cm}, \mathrm{~V}=19.0 \mathrm{~cm}$.


Figure 37. Mesohabitat J - midchannel;
Feb 2-5, 2002; $D=n / a, V=n / a$.


Figure 38. Mesohabitat J - right margin (code - 29);
Aug 27-30, 2002; $\mathrm{D}=65.7 \mathrm{~cm}, \mathrm{~V}=-5.0 \mathrm{~cm}$.


Figure 39. Allens Creek confluence (code - 30);
Aug 27-30, 2002; $\mathrm{D}=38.3 \mathrm{~cm}, \mathrm{~V}=7.3 \mathrm{~cm}$.


Figure 40. Allens Creek (code - 30);
Aug 27-30, 2002; $\mathrm{D}=38.3 \mathrm{~cm}, \mathrm{~V}=7.3 \mathrm{~cm}$.

Table 12. Documented occurrences of fish species near our Brazos River study reach (from Linam et. al 1994 and Winemiller et. al. 2000). * indicates species not collected during our study.

| Family | Species | Common Name |
| :---: | :---: | :---: |
| Amiidae | Amia calva | bowfin |
| Aphredoderidae | Aphredodreus sayanus | pirate perch |
| Atherinidae | Labidesthes sicculus Menidia beryllina | brook silverside inland silverside |
| Catostomidae | Carpoides carpio Cycleptus elongatus * Ictiobus bubalus Minytrema melanops * | river carpsucker <br> blue sucker <br> smallmouth buffalo <br> spotted sucker |
| Centrarchidae | Elassoma zonatum * <br> Lepomis cyanellus <br> Lepomis gulosus <br> Lepomis humilis <br> Lepomis macrochirus <br> Lepomis megalotis <br> Lepomis microlophus <br> Lepomis punctatus <br> Micropterus punctulatus <br> Micropterus salmoides <br> Pomoxis annularis <br> Pomoxis nigromaculatus * | banded pygmy sunfish green sunfish warmouth orangespotted sunfish bluegill sunfish longear sunfish redear sunfish spotted sunfish spotted bass largemouth bass white crappie black crappie |
| Cichlidae | Oreochromis aureus * | blue tilapia |
| Clupeidae | Alosa crysochloris Dorosoma cepedianum Dorosoma petenense | skipjack herring gizzard shad threadfin shad |
| Cyprinidae | Cyprinella lutrensis <br> Cyprinella venusta <br> Cyprinus carpio <br> Hybognathus nuchalis * <br> Lythrurus fumeus <br> Macrhybopsis aestivalis <br> Machrybopsis storeriana <br> Notemigonus crysoleucas * <br> Notropis buchanani <br> Notropis buccula * <br> Notropis oxyrhynchus <br> Notropis shumardi <br> Opsopoeodus emiliae <br> Pimephales vigilax | red shiner <br> blacktail shiner <br> common carp <br> Mississippi silvery minnow <br> ribbon shiner <br> speckled chub <br> silver chub <br> golden shiner <br> ghost shiner <br> smalleye shiner <br> sharpnose shiner <br> silverband shiner <br> pugnose minnow <br> bullhead minnow |

Table 12. Documented occurrences of fish species near our Brazos River study reach (from Linam et. al 1994 and Winemiller et. al. 2000). * indicates species not collected during our study. (continued)

| Fundulidae | Fundulus notatus | blackstripe topminnow |
| :--- | :--- | :--- |
| Ictaluridae | Ameiurus melas * <br> Ameiurus natalis * <br> Ictalurus furcatus <br> Ictalurus punctatus <br> Noturus gyrinus <br> Pylodictis olivaris | black bullhead <br> yellow bullhead <br> blue catfish <br> channel catfish <br> tadpole madtom <br> flathead catfish |
|  | Lepidosteus oculatus <br> Lepidosteus osseus | spotted gar <br> longnose gar |
| Lepisosteidae | Mugil cephalus <br> Mugil curema * | striped mullet <br> white mullet |
| Mugilidae | Etheostoma chlorosomum * <br> Eercidae | bluntnose darter <br> slough darter |
|  | Percina caprodes * <br> Percina macrolepida * <br> Percina sciera * | logperch <br> bigscale logperch <br> dusky darter |
| Poeciliidae | Gambusia affinis | mosquitofish |
| Sciaenidae | Aplodinotus grunniens | freshwater drum |

### 4.4 Index of Biological Integrity

Scores for the 10 IBI metrics of the seined collections are reported in Table 13. Our study reach rated good (score: 63) for September 2001 and excellent (score: 69) across our six collections. When considering all sampling gears our study reach rated excellent (score: 71) in both September and overall collections (Table 14). Our study reach scored consistently higher than the scores for seined collections at six sites (22 to 63), and seine and electrofish collections at three of six sites (44 to 53) calculated by Winemiller and Gelwick (1999). Differences in scores and categorical rankings between the two studies may be attributed to differences in the total area sampled. Winemiller and Gelwick (1999) sampled between 25-200 m of river length per site whereas our site encompassed over 4950 m , increasing the likelihood of capturing species of low densities or abundances.

Table 13. IBI ranks and scores of seine captures during September 2001 and overall collections.

## Scoring Criteria

September 2001 Collections
Overall Collections
Value Score
Value
Score

Species Richness and Composition Metrics:

| \# of native <br> species | 26 | 10 | 38 | 10 |
| :--- | :---: | :---: | :---: | :---: |
| \# of darter <br> species | 0 | 0 | 2 | 5 |
| \# of sunfish <br> species | 5 | 10 | 6 | 10 |
| \# of sucker <br> species | 1 | 10 | 2 | 10 |
| \# of intolerant <br> species | 7 | 7 | 9 | 10 |
| \% tolerant <br> species | 87.15 | 5 | 4.12 | 5 |

Trophic Function Metrics:

| \% omnivores | 76.03 | 7 | 87.35 | 5 |
| :--- | :---: | :---: | :---: | :---: |
| \% invertivores | 23.02 | 7 | 12.44 | 5 |
| \% carnivores | 0.70 | 2 | 0.18 | 2 |
| Totals: |  | 63 (good) |  | 69 (excellent) |

Table 14. IBI ranks and scores of total captures during September 2001 and overall collections.

## Scoring Criteria

September 2001 Collections
Overall Collections


Table 15. Comparison of IBI scores for mainstem reaches of the lower Brazos River.
Source Scoring Rating

Winemiller and Gelwick
(seine only):
Range: 22-63
Poor - Good
September-October 1998
Winemiller and Gelwick
(seine and electrofish):
Range: 44-53
Fair - Good
September-October 1998
TWDB (seine only):
September 2001
63
Overall
69
Good
Excellent
TWDB (total collections):
September 2001
71
Excellent
Overall
71
Excellent

### 4.5 Fish Species Indicators

Of the common species, bullhead minnow had the highest indicator value of pools but was not-significant ( $P>0.05$; Table 16). Red shiner and striped mullet had the highest values for runs, but were also not significant. Riffles were poorly differentiated by fishes of any species. Ghost shiner, silverband shiner, and mosquitofish had the highest indicator values of the tributary confluence habitat, with mosquitofish being the only significant indicator species. Results of an indicator species analysis conducted for all captured species is reported in Table 17.

Table 16. Indicator values for common fishes (abundance $>1 \%$ ) based on relative abundance and frequency of occurrence in Brazos River mesohabitats. $P$ is the proportion of Monte Carlo randomized trials (1000) with indicator values equal to or exceeding the observed indicator value. Bold numbers indicate the value that is highest for each species.

|  |  | Mesohabitat |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Species | $P$ | Pool | Run | Riffle | Tributary <br> Confluence |
|  |  |  |  |  |  |
| Red shiner | 0.412 | 16 | $\mathbf{5 1}$ | 14 | 19 |
| Ghost shiner | 0.612 | 14 | 12 | 0 | $\mathbf{4 5}$ |
| Silverband shiner | 0.179 | 28 | 14 | 3 | $\mathbf{5 5}$ |
| Bullhead minnow | 0.064 | $\mathbf{5 9}$ | 17 | 1 | 23 |
| Striped mullet | 0.686 | 20 | $\mathbf{2 2}$ | 0 | 2 |
| Mosquitofish | 0.003 | 5 | 6 | 0 | $\mathbf{8 7}$ |

Table 17. Indicator values for all fishes based on relative abundance and frequency of occurrence in Brazos River mesohabitats. $P$ is the proportion of Monte Carlo randomized trials (1000) with indicator values equal to or exceeding the observed indicator value. Bold numbers indicate the value that is highest for each species.

| Species | $P$ | Pool | Mesohabitat |  | Tributary Confluence |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Run | Riffle |  |
| Bowfin | 0.999 | 0 | 17 | 0 | 0 |
| Pirate perch | 0.999 | 0 | 17 | 0 | 0 |
| Brook silverside | 0.999 | 0 | 0 | 0 | 17 |
| Inland silverside | 0.002 | 1 | 0 | 0 | 80 |
| River carpsucker | 0.164 | 29 | 37 | 0 | 10 |
| Smallmouth buffalo | 0.707 | 19 | 19 | 0 | 11 |
| Green sunfish | 0.999 | 1 | 0 | 0 | 16 |
| Warmouth | 0.207 | 0 | 33 | 0 | 0 |
| Orangespotted sunfish | 0.067 | 1 | 1 | 0 | 46 |
| Bluegill sunfish | 0.006 | 0 | 1 | 0 | 64 |
| Longear sunfish | 0.871 | 2 | 20 | 0 | 10 |
| Redear sunfish | 0.999 | 2 | 0 | 0 | 15 |
| Spotted sunfish | 0.999 | 17 | 0 | 0 | 0 |
| Hybrid sunfish | 0.999 | 0 | 17 | 0 | 0 |
| Juvenile sunfish | 0.294 | 0 | 4 | 0 | 29 |
| Spotted bass | 0.999 | 0 | 1 | 0 | 15 |
| Largemouth bass | 0.999 | 0 | 0 | 0 | 17 |
| White crappie | 0.999 | 0 | 7 | 0 | 9 |
| Skipjack herring | 0.999 | 17 | 0 | 0 | 0 |
| Gizzard shad | 0.222 | 12 | 25 | 0 | 40 |
| Threadfin shad | 0.041 | 3 | 8 | 0 | 66 |
| Red shiner | 0.828 | 21 | 30 | 22 | 27 |
| Blacktail shiner | 0.122 | 7 | 30 | 0 | 0 |
| Common carp | 0.999 | 17 | 0 | 0 | 0 |
| Ribbon shiner | 0.388 | 5 | 23 | 0 | 0 |
| Speckled chub | 0.483 | 31 | 13 | 26 | 1 |
| Silver chub | 0.589 | 22 | 5 | 4 | 0 |
| Ghost shiner | 0.492 | 12 | 12 | 0 | 44 |
| Sharpnose shiner | 0.999 | 0 | 0 | 0 | 17 |
| Silverband shiner | 0.059 | 17 | 8 | 3 | 71 |
| Pugnose minnow | 0.999 | 8 | 8 | 0 | 0 |
| Bullhead minnow | 0.259 | 24 | 19 | 3 | 54 |
| Blackstripe topminnow | 0.177 | 0 | 0 | 0 | 33 |
| Blue catfish | 0.861 | 17 | 12 | 0 | 17 |
| Channel catfish | 0.766 | 24 | 19 | 1 | 21 |
| Tadpole madtom | 0.999 | 3 | 0 | 0 | 14 |
| Flathead catfish | 0.046 | 49 | 0 | 0 | 7 |
| Spotted gar | 0.327 | 7 | 13 | 0 | 36 |
| Longnose gar | 0.215 | 38 | 35 | 0 | 10 |
| Striped mullet | 0.897 | 21 | 19 | 0 | 3 |
| Slough darter | 0.999 | 0 | 0 | 0 | 17 |
| Mosquitofish | 0.005 | 2 | 7 | 0 | 89 |
| Freshwater drum | 0.048 | 50 | 0 | 0 | 0 |

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